Contents lists available at ScienceDirect





Journal of Hazardous Materials

journal homepage: www.elsevier.com/locate/jhazmat

Aerobic granules with inhibitory strains and role of extracellular polymeric substances

Sunil S. Adav^{a, 1}, Duu-Jong Lee^{a,*}, Juin-Yih Lai^b

^a Department of Chemical Engineering, National Taiwan University, Taipei 10617, Taiwan

^b Center of Membrane Technology, Department of Chemical Engineering, Chung Yuan Christian University, Chungli, Taiwan

ARTICLE INFO

Article history: Received 27 May 2009 Received in revised form 11 September 2009 Accepted 14 September 2009 Available online 20 September 2009

Keywords: Aerobic granules FPS Inhibition Physical isolation Adsorption

ABSTRACT

Microorganisms compete with other species by secreting antimicrobial compounds. The compact structure of aerobic granules was generally assumed to provide spatial isolation, resulting in the co-occurrence of diverse strains that have similar or dissimilar functions. No studies have investigated whether stable, mature aerobic granules can be formed with two mutually inhibitory strains. The strain Acinetobacter sp. 18 competes with Bacillus sphaericus 15 in a well-mixed environment, but can form stable and mature granules at 400 mg L⁻¹ phenol by repeatedly replenishing fresh medium in a sequencing batch reactor. The supernatants collected from the I8 medium in its exponential-growth phase or from the I5+I8 medium cultivated for 12 or 24 h significantly inhibited I5 growth. Addition of tightly bound extracellular polymeric substances (TBEPS) or loosely bound extracellular polymeric substances (LBEPS) extracted from I5 + I8 granules effectively suppressed the inhibitory effects of I8 on I5. The TBEPS or LBEPS physically separate strain 15 from 18 in the granule, and effectively adsorb the inhibitory substance(s) in the suspension.

© 2009 Elsevier B.V. All rights reserved.

1. Introduction

The application of aerobic granular sludge is considered as a promising biotechnology in wastewater treatment [1–3]. The first patent for the use of aerobic granules was granted to Heijnen and van Loosdrecht [4]. Aerobic granules have dense and strong structures, good settleability, high biomass retention, and high tolerance to medium toxicity [5,6]. Although granulation process has been characterized [7–10], little is known about the microbial interactions and adaptive mechanism for microbial survival within aerobic granules. The two populations associate tightly under some conditions or location indicating the beneficial association. In contrast, the clusters of population remain at a distance from each other in the community within aerobic granules when the productions of antagonist or competition exist. During such conditions, microbial cells get entrapped within the matrix of extracellular polymeric substances (EPS) secreted by themselves, a universal survival strategy adopted by microbes. EPS are metabolic products that are major components of activated sludge flocs, biofilms and microbial granules [11,12] and their layer forms a protective shield for aerobic granule cells against harsh external environments [13].

Aerobic granules are composed of numerous microbial strains [14]. Jiang et al. [15] isolated 10 strains from phenol-degrading granules that were either good phenol reducers or good flocculators. Jiang et al. [16] demonstrated that the two functionally dissimilar isolates, Propioniferax-like PG-02 (fast-growing strain in phenol) and Comamonas sp. PG-08 (strong coagulator) cannot coexist at 250 mg L^{-1} phenol in a completely mixed environment, but can co-exist in a spatially heterogeneous environment. Jiang et al. [17] determined that two functionally similar strains, Pandoraea sp. PG-01 and Propioniferax-like PG-02, are fast-growing strains in phenol; however, they cannot co-exist in a well-mixed medium due to mutual competition. Zhou et al. [18] and Treves et al. [19] indicated that competitively inferior strains can co-exist when physical isolation is provided. Jiang et al. [17] demonstrated that two functionally similar strains, Pandoraea sp. PG-01 and Rhodococcus erythropolis PG-03 obtained from their phenol-fed aerobic granules, cannot coexist in a well-mixed medium due to mutual competition, but can co-exist in the spatially heterogeneous structure of aerobic granules. However, these studies did not investigate whether the strains studied are mutually inhibitory.

Adav and Lee [20] isolated nine strains (I1-I9) from their aerobic phenol-degrading granules; strains Bacillus thuringiensis I2, Acinetobacter calcoaceticus I6 and Acinetobacter sp. I8 have high auto-aggregation capabilities and can form single-strain granules

^{*} Corresponding author. Tel.: +886 2 23625632; fax: +886 2 23623040. E-mail addresses: adavs@rediffmail.com (S.S. Adav), djlee@ntu.edu.tw, djlee@ccms.ntu.edu.tw (D.-J. Lee), jylai@cycu.edu.tw (J.-Y. Lai).

¹ Tel.: +886 2 23625632; fax: +886 2 23623040.

^{0304-3894/\$ -} see front matter © 2009 Elsevier B.V. All rights reserved. doi:10.1016/j.jhazmat.2009.09.070

[21]. These authors also determined that the strains *Acinetobacter* sp. 18 and *Bacillus sphaericus* 15 were mutually inhibitory. Zhou et al. [18] and Treves et al. [19] claimed that EPS physically isolate strains from mutual competition and/or reduce local mass transfer of toxins to cells.

No studies have investigated whether stable, mature granules can be formed using inhibitory strains. This study tested two strains, one inhibits the growth of another, in homogeneous (well mixed) and heterogeneous (still) media; the two strains were cultivated to form stable and mature aerobic granules. This work provides experimental evidence that the EPS extracted from the I5+18 granules effectively eliminated the secreted inhibitory substance(s). The EPS provided physical isolation for strain I5 from I8, and eliminated the secreted inhibitory substance(s).

2. Materials and methods

2.1. Strains and medium

This study used bacterial strains *Acinetobacter* sp. I8 and *B. sphaericus* I5, which were isolated previously from phenoldegrading granules [20]. Strain I8 is a Gram-negative bacterium that has a short rod shape, and strain I5 is a Gram-positive bacterium with a rod-shaped morphology. These two strains have high phenol-degrading capability. The composition of the MP medium used in this work was (in mg L⁻¹): 1000, (NH₄)₂SO₄; 200, MgCl₂; 100, NaCl; 20, FeCl₃; 10, CaCl₂; 1350 KH₂PO₄ and 1650 K₂HPO₄ (pH 6.8). The micronutrients were (mg L⁻¹): 50, H₃BO₃; 50, ZnCl₂; 30, CuCl₂; 50, MnSO₄·H₂O; 50, (NH₄)Mo₇O₂₄·4H₂O; 50, AlCl₃; 50, CoCl₂·6H₂O and 50, NiCl₂.

2.2. Granules cultivation and EPS extraction

Aerobic granules were cultivated in column-type sequential batch reactors (SBR) 6 cm in diameter and 120 cm in height. These reactors seeded with 2L of 15 + 18 and fed with sterilized MP medium at pH 6.8 ± 0.2 with 250 mg L^{-1} phenol as the sole carbon source. Fine air bubbles at a flow rate of 3 Lmin^{-1} were supplied at the reactor bottom, and the air outlet was immersed in sterilized water. The column was operated at cycle time of 6 h (5 min settling, 5 min filling, 5 min effluent withdrawal, and 5 h and 45 min of aeration). The volumetric exchange ratio 50% was applied by discharging the effluent above 50 cm from the reactor bottom followed by replenishing the reactor with the same volume of fresh sterilized medium in each cycle. The reactor was operated for 50 days.

The EPS were extracted from cultured I5 + I8 aerobic granules as described in [22]. In brief, the samples were washed with water, and loosely bound EPS (LBPES) were obtained by centrifugation at $5000 \times g$ for 10 min. The residues were resuspended to their original volume using saline solution (0.05% NaCl), and were extracted again using low-frequency ultrasound at 20W for 5 min in an ice bath. Following ultrasonication, suspensions were collected by centrifugation and filtered through a 0.2 µm filter (Advanced Microdevices, Ambala Cantt, India). The EPS in the collected filtrate were considered the tightly bound EPS (TBEPS) of the sample. Via quantification of the 2-keto-3-deoxyoctonate (KDO) in the extract, the quantities of DNA in all extracted EPS samples were <0.2 mg g⁻¹ volatile suspended solids (VSS), indicating negligible contamination of the collected EPS by intracellular matter. The carbohydrate content in EPS was measured by the Anthrone method [23] with glucose as the standard. The protein and humic content in EPS was measured by the modified Lowry method [24] using bovine serum albumin and humic acid (Fluka, USA) as the respective standards.

The granules were collected, washed with phosphate buffered saline (PBS, pH 7.2) and fixed with 4% paraformaldehyde for 3 h at 4°C. The fixed granules were washed with PBS buffer and stained for β-polysaccharides by adding calcofluor white (fluorescent brightener 28, Sigma, USA) solution $(300 \text{ mg L}^{-1}, 100 \mu\text{L})$ for 30 min. The stained granule was washed twice with PBS to remove excess stain and hybridized for FISH as described by [25] with hybridization buffer containing $5 \text{ ng }\mu\text{L}^{-1}$ of each of the specific probes—Acinetobacter sp. (ATC CTC TCC CAT ACT CTA) and B. sphaericus (ATG AGA AAT TTG GAT TTT ATT)-labeled with FAM (green) and Cy3 (red). The granule was then embedded for cryosectioning in embedded medium (Shandon Cryomatrix, Pittsburgh, PA, USA). Embedded samples were frozen at -20 °C, after which 40µm sections were cut on a cryomicrotome and mounted onto a gelatin-coated (0.1% gelatin and 0.01% chromium potassium sulfate) microscopic slide and analyzed by confocal laser scanning microscopy (CLSM) (Leica TCS SP5, Confocal Spectral Microscope Imaging System, GmbH, Germany).

2.3. Homogeneous and heterogeneous tests

Reagent bottles (500 ml) containing 200 ml sterilized MP medium and 400 mg L⁻¹ phenol were utilized in tests. Equal quantities of strains I5 and I8 at their respective exponential-growth phases were inoculated and incubated at 35 °C and 150 rpm in a rotary shaker to produce a well-mixed (homogeneous) environment. Spatially heterogeneous tests were conducted following the procedures developed by Rainey and Travisano [26]. The experimental protocol was the same as in homogeneous environment tests, except that bottles were kept still and were mixed manually daily.

Suspension samples were collected from homogeneous environment tests and heterogeneous environment tests during their daily, manual mixing period. The bacteria in the collected samples were concentrated by centrifugation (8000 rpm at 4 °C), washed with 1× PBS and fixed with 4% paraformaldehyde for 30 min at 4 °C. The fixed samples were resuspended in 50% ethanol after washing with 1× PBS buffer. The resuspended cells were then hybridized using hybridization buffer (0.9 M NaCl, 20 mM Tris–HCl at pH 7.4, 0.01% sodium dodecyl sulfate and formamide) containing 5 ng μ L⁻¹ of the probes labeled with FAM and Cy3 probes, as stated in Section 2.2. The hybridized cells were imaged via CLSM and cell populations (%) were determined their respective fluorescent signals.

2.4. Inhibitory tests with extracted EPS

In total, 8 ml of strain 18 (OD=0.8) was incubated in 500 ml reagent bottles containing 200 ml sterilized MP medium with 400 mg L⁻¹ phenol at 35 °C in a well-mixed environment. The medium supernatant was collected separately during the exponential-growth phase (supernatant 18-E) and stationary phase (18-S) of 18 cells following centrifugation and filtration. Moreover, the supernatants of mixed cultures collected (Section 2.3) with 15 + 18 under a well-mixed environment were collected separately at 12 h (15 + 18-12 h) and 24 h (15 + 18-24 h) of incubation.

In total, 8 ml of strain 15 was incubated at 35 °C in 500 ml reagent bottles containing 100 ml sterilized MP medium with 400 mg L⁻¹ phenol, and 100 ml of one of the four collected supernatants (I8-E, I8-S, I5 + I8-12 h, and I5 + I8-24 h). Each set of bottles fed with a specific supernatant was further divided into three groups. The first group was fed with 150 ml of LBEPS from I5 + I8 granules (Section 2.2); the second group was fed with 150 ml of TBEPS from granules (Section 2.2); the third group without EPS was used as a control. All bottles were shaken at 150 rpm to generate a homogeneous environment.

2.5. Analytical methods

The dry weights of granules, VSS, SS, and the sludge volume index (SVI) in the suspension were measured according to Standard Methods [27]. The size of the granules was determined by a particle size analyzer (Mastersizer Series 2600; Malvern Instruments, Worcestershire, UK). Phenol concentrations in the reactor were determined by high-performance liquid chromatography (HPLC) equipped with a C18 column (Varian, Inc., CA, USA) at wavelength 276 nm. The mobile phase comprised of acetonitrile:water (300:700), 0.11 g heptane sulphonic acid, 0.29 g anhydrous sodium acetate, and 2.5 ml glacial acetic acid. The size exclusion chromatography system used for EPS analysis comprised a BETA 10 gradient pump (Ecom spol. s r. o., Prague, Czech Republic), a size exclusion TSK G3000SW_{XI} column (TOSOH Bioscience, Stuttgart, Germany), on-line SAPPHIRE 600 UV-VIS variable wavelength detector (Ecom spol. s. r. o., Prague, Czech Republic) and a CHF 100SA fraction collector (Advantec MFS, Inc., Dublin, CA, USA).

3. Results

3.1. Homogeneous and heterogeneous tests

The changes in populations of 15 and 18 cells over time in the homogeneous medium were monitored (Fig. 1a). In homogeneous medium, the number of 18 cells increased from $54.4 \pm 3.5\%$ initially to roughly $97 \pm 1.1\%$ after 24 h of incubation while 15 declined to 2.7% from its initial population of $45.5 \pm 3.5\%$ (Fig. 1a). In other



Fig. 1. Proportion of cell populations in different environment at 400 mg L⁻¹ phenol concentration. Cells were hybridized with specific probes and the percentage of each strain was calculated. (a) Homogeneous medium and (b) heterogeneous medium.



Fig. 2. Cell populations for strain 15 with MP medium and supernatants I8-E (supernatant A), I8-S (supernatant B), 15+18-12h (supernatant C) and 15+18-24h (supernatant D) (40:60, v/v). (a) With no added EPS; (b) with added TPEPS and (c) with added LBEPS.

words, strain 18 competes with strain 15 in the homogeneous medium. In the heterogeneous medium, conversely, the population of 15 decreased by 27–30% and stabilized after 24 h of incubation (Fig. 1b). The heterogeneous medium shielded strain 15 from strain 18. This experimental observation correlates with findings obtained by Jiang et al. [17], Zhou et al. [18], Treves et al. [19], indicating that competitive strains can co-exist when they are physically separated.

3.2. Inhibition test

The I5 strain grew well in the MP medium, reaching $7.7 \pm 0.5 \times 10^8$ CFU ml⁻¹ during cultivation for 16 h (control in Fig. 2a). With supernatant I8-E (supernatant collected at the exponential-growth phase of strain I8), I5+I8-12 h (supernatant collected during the homogeneous test with I5 + I8 after 12 h of testing) or I5+I8-24 h (supernatant collected during the homogeneous test with I5 + I8 after 24 h testing) added, the growth of I5 cells was inhibited significantly. For example, with I8-E added, the concentrations of I5 cells were $3.1 \pm 0.6 \times 10^8$ and $0.8 \pm 0.4 \times 10^8$ CFU ml⁻¹ after 16 and 120 h cultivation, respectively, accounting for 40% and 15% of the control respectively.

Adding TBEPS or LBEPS collected from I5 + 18 granules (Section 2.2) markedly decreased the inhibitory effects of I8-E, I5 + I8-12 h or I5 + I8-24 h on strain I5 (Fig. 2b and c). For example, when TBEPS were added, the quantities of I5 cells in I8-E were $5.2 \pm 0.5 \times 10^8$ and $4.1 \pm 0.4 \times 10^8$ CFU ml⁻¹ after 16 or 120 h cultivation, respectively, significantly higher than those without TBEPS added $(3.1 \pm 0.6 \times 10^8 \text{ and } 0.8 \pm 0.4 \times 10^8 \text{ CFU ml}^{-1}$, respectively).

3.3. Cultivated granules and extract EPS

In granule cultivation tests, small aggregates of I5 + I8 appeared within 1 day in the SBR. The granules grew to 0.5-0.7 mm in size in 14 days, and the corresponding SVI values decreased from 210 to 112 ml g^{-1} . On day 30, the granules had grown to a mean size of 1.6 mm and SVI values declined to 52.8 ml g^{-1} . Microscopic observations demonstrate that the formed granules had a nearly spherical shape and smooth surface. The strains in granules were *Acinetobacter* sp. 18 and *B. sphaericus* 15 via analysis of their corresponding 16S rRNA sequences in extracted DNA.

Chemical analysis reveals that the quantities of proteins, polysaccharides, and humic substances in the EPS extracted



Fig. 3. FISH–CLSM image of aerobic granule cross-sectioned at 360 μ m from top surface, stained for β -polysaccharides (blue: calcofluor white) and simultaneously hybridized with specific probes for strain 18 (green) and strain 15 (red). Bar represents 200 μ m. (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article.)

from 15+18 granules were 228.4 ± 21.8 , 145.8 ± 18.2 , and $109.2 \pm 8.9 \text{ mg g}^{-1}$ VSS, respectively. Retention times of SEC chromatograms of LBEPS were roughly 5 and 9.1 min, whereas that of TBEPS was at 10.9 min (single peak). The apparent molecular weights (AMWs) of LBEPS and TBEPS were estimated at >20,000 and 8000 Da, and about 2000 Da, respectively. The residue (pellet) after EPS extraction has a broad range of AMWs, ranging from <100 to >20,000 Da.

Fig. 3 shows the FISH–CLSM image of an 15+18 aerobic granule, cross-sectioned at 360 μ m from the top surface. The β -polysaccharide was distributed in the entire granule structure forming a backbone. The β -polysaccharides (blue: calcofluor white) physically separated strain 18 (green) and strain 15 (red).

4. Discussion

4.1. Inhibition of I8 on I5 in homogeneous and heterogeneous tests

The I5 and I8 cells cannot co-exist in a homogeneous medium, based on the inhibitory effects of I8 on I5 in plate tests [20]. The supernatant I8-E rather than I8-S inhibited I5 growth (Fig. 2a). Hence, the inhibitory substance(s) were primarily secreted/generated with phenol metabolism (intermediates of phenol metabolism) by I8 during its exponential-growth phase, rather than during its stationary phase.

The I5+I8-12 h and I5+I8-24 h supernatants inhibited I5 growth, demonstrating that I8 secreted and released inhibitory substance(s) into surrounding liquid when I5 is present. In the homogeneous medium, the inhibitory substance(s) efficiently reached I5 via the mixing current. When I5 and I8 were kept still in a bottle (heterogeneous tests), the I5 cells co-existed with I8 from 24 h and beyond (Fig. 1b) as the inhibitory substance(s) secreted by I8 that effectively inhibit I5 cells diffuse slowly or due to low concentration of inhibitory substance.

4.2. Granules with mutually inhibitory strains

When I5 + I8 are co-cultivated in an SBR to form granules, likely due to dilution via frequent replenishment of fresh MP medium during SBR operation, the inhibitory substance(s) secreted by I8 did not inhibit I5's growth.

As revealed by Fig. 3, the EPS physically separated strain I5 and I8, as demonstrated by Zhou et al. [18] and Treves et al. [19]. Prolonged tests indicate that the I5+I8 granules were stable over a subsequent 1-month test in MP+phenol medium in an SBR. The I5+I8 granules were a stable eco-system for the co-existence of two inhibitory strains.

4.3. Role of EPS on inhibitory effects

Adding extracted TBEPS or LBEPS generally eliminate the inhibition effects of supernatants I8-E, I5 + I8-12 h or I5 + I8-24 h on I5 cells (Fig. 2b and c). Hence, the EPS shielded the I5 strain from inhibitory substance(s) secreted by I8 or generated during phenol metabolism.

A test with 200 ml sterilized MP medium and 400 mg L⁻¹ phenol, 2 ml each for strains 15 and 18 at their respective exponentialgrowth phase, and 150 ml TBEPS extracted from 15+18 granules was conducted in reagent bottles at 35 °C and 150 rpm. Fig. 4 shows the FISH–CLSM images of strains 18 and 15 after 50 h incubation. The quantities of both 15 and 18 were considerable in the medium, implying that 18 only minimally inhibited 15. The 15 cells remained in a dispersed state (Fig. 4); hence, the TBEPS added in this test do not provide a physically isolated environment for 15 cells, as suggested by Zhou et al. [18] and Treves et al. [19], or provide



Fig. 4. FISH–CLSM image of strains 18 and 15 in homogeneous MP medium and extracted EPS from aerobic granules (40:60, v/v) after 50 h. In situ hybridization was performed simultaneously with specific probes labeled with Cy3 and FAM (red-strain 18 and green-strain 15). Bar represents 10 μ m. (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article.)

mass transfer resistance to reduce local concentrations of toxins in suspension, as argued by Chiu et al. [13]. Adsorption of inhibitory substance(s) on EPS molecules may be a mechanism for the noted shield effect by the EPS. However, although the extracted TBEPS or LBEPS have extremely different AWM distributions, both effectively "protect" 15 from 18, demonstrating that the hydrophobic interaction or "van der Walls interaction" between the functional groups on the EPS, and inhibitory substances did not likely correspond to the proposed adsorption mechanisms.

5. Conclusions

Microorganisms inhibit the outgrowth of other species by secreting antimicrobial compounds. This inhibition due to antimicrobial substance secreted by one species can be overcome by strategies such as cell immobilization to protect other microbial cells. In the compact structured aerobic sludge granules, EPS provide spatial isolation for the microbial strains having similar or dissimilar functions. In homogeneous environment, strain *Acinetobacter* sp. 18 inhibits *B. sphaericus* 15, while they co-exits in aerobic granules. Thus, role of EPS is proposed to effectively adsorb the inhibitory substance(s) and provide spatial isolation environment for the strain.

Acknowledgment

This project is supported by National Natural Science Foundation of China (50876024).

References

- J.J. Beun, A. Hendriks, M.C.M. van Loosdrecht, E. Morgenroth, P.A. Wilderer, J.J. Heijnen, Aerobic granulation in a sequencing batch reactor, Water Res. 33 (1999) 2283–2290.
- [2] E. Morgenroth, T. Sherden, M.C.M. van Loosdrecht, J.J. Heijnen, P.A. Wilderer, Aerobic granular sludge in a sequencing batch reactor, Water Res. 31 (1997) 3191–3194.
- [3] S. Yi, W.Q. Zhuang, B. Wu, S.T.L. Tay, J.H. Tay, Biodegradation of p-nitrophenol by aerobic granules in a sequencing batch reactor, Environ. Sci. Technol. 40 (2006) 2396–2401.
- [4] J.J. Heijnen, M.C.M. van Loosdrecht, Method for acquiring grain-shaped growth of a microorganism in a reactor. European Patent EP0826639 (1998).
- [5] J.H. Tay, Q.S. Liu, Y. Liu, The effects of shear force on the formation, structure and metabolism of aerobic granules, Appl. Microbiol. Biotechnol. 57 (2001) 227–233.
- [6] K.Z. Su, H.Q. Yu, Formation and characterization of aerobic granules in a sequencing batch reactor treating soybean-processing wastewater, Environ. Sci. Technol. 39 (2005) 2818–2828.
- [7] Y. Liu, J.H. Tay, State of the art of biogranulation technology for wastewater treatment, Biotechnol. Adv. 22 (2004) 533–563.
- [8] N. Maximova, O. Dahl, Environmental implications of aggregation phenomena: current understanding, Curr. Opin. Colloid Interface Sci. 11 (2006) 246– 266.
- [9] M.K. de Kreuk, N. Kishida, M.C.M. van Loosdrecht, Aerobic granular sludge-state of the art, Water Sci. Technol. 45 (2007) 81-85.
- [10] S.S. Adav, D.J. Lee, K.Y. Show, J.H. Tay, Aerobic granular sludge: recent advances, Biotechnol. Adv. 26 (2008) 411–423.
- [11] J.A.S. Goodwin, C.F. Forster, A further examination into the composition of activated sludge surfaces in relation to settlement characteristics, Water Res. 19 (1985) 527–533.
- [12] Y. Liu, H.H.P. Fang, Influences of extracellular polymeric substances (EPS) on flocculation, settling and dewatering of activated sludge, Crit. Rev. Environ. Sci. Technol. 33 (2003) 237–273.
- [13] Z.C. Chiu, M.Y. Chen, D.J. Lee, S.T.L. Tay, J.H. Tay, K.Y. Show, Diffusivity of oxygen in aerobic granules, Biotechnol. Bioeng. 94 (2006) 505–513.
- [14] J. Xavier, M.K. de Kreuk, C. Picioreanu, M.C.M. van Loosdrecht, Multi-scale individual-based model of microbial and bioconversion dynamics in aerobic granular sludge, Environ. Sci. Technol. 41 (2007) 6410–6417.
- [15] H.L. Jiang, A.M. Maszenan, J.H. Tay, Bioaugmentation and coexistence of two functionally similar bacterial strains in aerobic granules, Appl. Microbiol. Biotechnol. 75 (2007) 1191–1200.
- [16] H.L. Jiang, J.H. Tay, A.M. Maszenan, S.T.L. Tay, Bacterial diversity and function of aerobic granules engineered in a sequencing batch reactor for phenol degradation, Appl. Environ. Microbiol. 70 (2004) 6767– 6775.
- [17] H.L. Jiang, J.H. Tay, A.M. Maszenan, S.T.L. Tay, Enhanced phenol biodegradation and aerobic granulation by two coaggregating bacterial strains, Environ. Sci. Technol. 40 (2006) 6137–6142.
- [18] J.Z. Zhou, B.C. Xia, D.S. Treves, L.Y. Wu, T.L. Marsh, R.V. O'Neill, A.V. Palumbo, J.M. Tiedje, Spatial and resource factors influencing high microbial diversity in soil, Appl. Environ. Microbiol. 68 (2002) 326–334.
- [19] D.S. Treves, B. Xia, J. Zhou, J.M. Tiedje, A two-species test of the hypothesis that spatial isolation influences microbial diversity in soil, Microbiol. Ecol. 45 (2003) 20–28.
- [20] S.S. Adav, D.J. Lee, Physiological characterization and interactions of isolates in phenol-degrading aerobic granules, Appl. Microbiol. Biotechnol. 78 (2008) 899–905.
- [21] S.S. Adav, D.J. Lee, Single-culture aerobic granules with Acinetobacter calcoaceticus, Appl. Microbiol. Biotechnol. 78 (2008) 551–557.
- [22] S.S. Adav, D.J. Lee, Extraction of extracellular polymeric substances from aerobic granules with compact interior structure, J. Hazard. Mater. 154 (2008) 1120–1126.
- [23] A.F. Gaudy, Colorimetric determination of protein and carbohydrate, Ind. Water Wastes 7 (1962) 17–22.
- [24] B. Frolund, K. Keiding, P.H. Nielsen, Enzymatic activity in the activated sludge flocs matrix, Appl. Microbiol. Biotechnol. 43 (1995) 755–761.
- [25] R.I. Amann, B.J. Binder, R.J. Olson, S.W. Chisholm, R. Devereux, D.A. Stahl, Combination of 16S rRNA-targeted oligonucleotide probes with flow cytometry for analyzing mixed microbial populations, Appl. Environ. Microbiol. 56 (1990) 1919–1925.
- [26] P.B. Rainey, M. Travisano, Adaptive radiation in a heterogeneous environment, Nature 394 (1998) 69-72.
- [27] APHA, The Standard Methods for the Examination of Water and Wastewater, 20th ed., American Public Health Association, Washington, DC, USA, 1998.